

## Research Article

# Application of Central Composite Design in the Adsorption of Ca(II) on Metakaolin Zeolite

Upenyu Guyo, Lycenter Yard Phiri, and Fidelis Chigondo

Department of Chemical Technology, Midlands State University, Private Bag 9055, Senga, Gweru, Zimbabwe

Correspondence should be addressed to Fidelis Chigondo; [chigondofj@gmail.com](mailto:chigondofj@gmail.com)

Received 15 May 2017; Revised 30 June 2017; Accepted 11 July 2017; Published 10 August 2017

Academic Editor: Ana Moldes

Copyright © 2017 Upenyu Guyo et al. This is an open access article distributed under the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original work is properly cited.

Metakaolin zeolite-A was synthesized from thermally activated kaolin clay and characterized by Fourier Transform Infrared Spectroscopy and X-Ray Diffraction Spectroscopy. The effects of pH (2–10), contact time (10–180 min), initial concentration (5–120 mgL<sup>-1</sup>), and dosage (0.1–2 g) and their interactions were investigated using response surface methodology following a central composite design. Optimum removal (87.70%) was obtained at pH 6, contact time 180 min, initial concentration 40.0 mgL<sup>-1</sup>, and adsorbent dosage 1.0 g by Excel Solver using the GRG solving method. The adsorption data fitted best to the Langmuir model with correlation coefficient  $R^2 = 0.993$  and Chi-square value  $\chi^2 = 4.76$ . The Freundlich isotherm gave a correlation coefficient  $R^2 = 0.933$  and  $\chi^2 = 37.91$ . The adsorption process followed the pseudo-second-order model. The calculated thermodynamic parameters showed that the adsorption process was endothermic and not thermodynamically spontaneous. The studied zeolite-A can therefore be used as a promising adsorbent for the removal of Ca(II) ions from aqueous solutions.

## 1. Introduction

Zeolites are crystalline, microporous, hydrated aluminosilicates formed by the sharing of oxygen atoms in the framework of aluminum and silicon tetrahedrals [1]. Synthetic zeolites are used commercially more often than natural zeolites due to the purity of their crystalline products, the uniformity of particles, the greater thermal stability, and the fact that they can be engineered with a wide variety of properties and pore sizes [2]. Zeolite synthesis involves the hydrothermal crystallization of aluminosilicate gels in the presence of alkali hydroxides [3]. The type of zeolite is affected by composition of the reaction mixture, the nature of the reactants and their pretreatments, temperature of the process, reaction time, and pH of the reaction mixture. The main problems in zeolite application are availability and cost of raw materials, specifically the silica source [4]. The use of commercial silica in zeolite synthesis is affected by variations in reactivity and selectivity apart from being expensive [5]. Cheaper raw materials such as clay minerals could be an alternative starting material for zeolite synthesis. Kaolin is a cheap clay mineral whose deposits are found vastly around mining areas in Zimbabwe. Many researchers have reported the

synthesis of zeolites from kaolin under different conditions [6–11]. Kaolin is not stable under severe alkaline conditions and as such it is desirable to convert it into a more reactive form (metakaolin) by calcination at elevated temperatures. Because of their unique porous properties, zeolites could be applied in the removal of hardness in domestic water which is mainly due to calcium and magnesium ions. Water supplies with total hardness of more than 500 mg/L are not suitable for domestic consumption [12]. Hard water is responsible for the formation of deposits in boilers, cooling towers, water pipes, and household facilities, as well as diverse influences on the cleaning performance of detergents. Traditionally, water purification plants used lime and soda ash for the removal of hardness. One of the main drawbacks of this process is the generation of large amounts of liquid sludge as well as the need for recarbonation of the treated water [13]. The use of polyphosphates in removing water hardness is also practiced but results in massive discharge of phosphates in water bodies which lead to eutrophication [14].

Various other methods including membrane precipitation, phytoextraction, flocculation, solvent extraction, ultrafiltration, reverse osmosis, electrodialysis, and adsorption have been studied for the removal of a wide variety of cations

from water streams including magnesium and calcium [15–17]. Most of these are either inefficient or expensive and result in the generation of large amounts of sludge. Activated carbon is considered to be a particularly competitive and effective process for the removal of cations but has been hampered by the high costs associated with production and regeneration of spent carbon. The use of low-cost ion exchangers such as zeolites is an attractive method for the removal of metals from aqueous solutions. In this work, we report on the synthesis of zeolite-A from Zimbabwean kaolin and its application in the sorption of calcium ions from aqueous solutions. All the process independent variables have been optimized collectively by applying response surface methodology (RSM), a mathematical statistical technique which optimizes the process and evaluates the relationship between a set of independent variables [18–20]. Moreover, it evaluates the relative significance of the various independent variables even in the presence of complex factor-factor interactions [21]. The adsorbent was characterized using Fourier Transform Infrared (FTIR) Spectroscopy and XRD (X-Ray Diffraction) Spectroscopy.

## 2. Materials and Methods

**2.1. Chemicals and Solutions.** All chemicals used in this work were obtained from Skylabs (Zimbabwe) and were used as received without further purification. A pH meter (Az-8601, China) was employed for pH measurement and 0.1 M HCl and 0.1 M NaOH were used for pH adjustment.  $\text{CaCl}_2$  was used for the preparation of stock solution using double deionized water. All glassware used was soaked in 10% (v/v) nitric acid solution for 24 h and cleaned repeatedly with double deionized water.

**2.2. Sample Collection and Pretreatment.** The kaolin clay used in this study was collected from Chegutu, Zimbabwe. The kaolin clay sample was pulverized and sieved through a 90  $\mu\text{m}$  sieve. The pulverized kaolin clay material was digested in 6 M HCl to remove iron and washed with distilled water followed by drying overnight at 120°C. The treated kaolin clay was exposed to 900°C for 1 h where its structure was destroyed [22]. This removed all incorporated hydrocarbons and was dehydroxylated to form an activated amorphous material called metakaolin, a material which accepts and readily exchanges sodium as a guest in its structure.

**2.3. Zeolitization Using the Hydrothermal Transformation Method.** The zeolite gel was prepared by dissolving metakaolin in sodium hydroxide solutions [23]. Metakaolin (20 g) was refluxed and stirred in 100 mL of sodium hydroxide (1, 3, and 5 M) at a temperature of 100°C at specified reaction times (1, 2, and 4 h) [5]. The metakaolin-sodium hydroxide slurries were filtered by suction filtration and washed several times with deionized water to remove excess unreacted sodium hydroxide. The reaction products were dried at a temperature of 120°C in the oven for 24 h. Triplicate samples were prepared for each NaOH concentration and reaction time.

**2.4. Characterization of the Zeolites.** The Fourier Transform Infrared (FTIR) Spectroscopy spectra of the metakaolin zeolite (MZ) were obtained using a Thermo Fisher Scientific spectrometer (Nicolet 6700, USA) in the scanning range 4000–400  $\text{cm}^{-1}$ . A Bruker D8 Advance X-ray diffractometer (Germany) using nickel filtered  $\text{Cu K}\alpha$  radiation (40 kV, 40 mA) was used for identification of the structural phases.

**2.5. Adsorption Batch Studies.** All experiments were carried out in batch mode in 250 mL conical flasks. Several experimental parameters (solution pH: 2–10; contact time: 10–180 min; adsorbent dosage: 0.1–2.0 g; and initial metal ion concentration: 5–120 mg/L) were investigated following a central composite design (CCD). A predetermined adsorbent mass was added to 100 mL solution in a conical flask containing  $\text{Ca}^{2+}$  cations at a concentration according to CCD. The mixture was agitated on a shaker at 150 rpm for a specified time at room temperature. After agitation, the solid matrix was separated from the rest of the solution by filtration using Whatman number 42 filter paper followed by centrifugation to get the supernatant solution. The determination of  $\text{Ca}^{2+}$  ions in aqueous solution was done by complexometric titration using standardized 0.1 M EDTA (ethylenediaminetetraacetic acid) in the presence of Eriochrome Black T as an indicator. Typically, 100 mL of each supernatant solution was pipetted into a 250 mL conical flask, 4 mL of 1.0 M  $\text{NH}_3$  solution and 2–3 drops of Eriochrome Black T were added, and the mixture was titrated with 0.1 M EDTA solution until the color changed from pink to blue [12].

The percentage removal ( $R$ ) of Ca(II) ions from aqueous solutions was determined using the following equation:

$$R = \frac{(C_o - C_e)}{C_o} \times 100\%. \quad (1)$$

The equilibrium adsorption amount ( $q_e$ ) was calculated according to the mass balance equation:

$$q_e = \frac{(C_o - C_e)V}{M}, \quad (2)$$

where  $q_e$  is the equilibrium amount ( $\text{mg g}^{-1}$ ) of Ca(II) ions adsorbed per gram of metakaolin zeolite adsorbent,  $C_o$  and  $C_e$  are the Ca(II) concentrations ( $\text{mg L}^{-1}$ ) in the solution initially and after adsorption, respectively,  $V$  is the volume (L) of the solution, and  $M$  is the mass (g) of the adsorbent used in the experiment.

The experimental domain used is presented in Table 1. Preliminary experiments were done to obtain the extreme levels of the independent input variables. Design Expert Version 5 (Stat-Ease, USA) was used to generate the CCD experiments, optimize the levels of the independent variables, and evaluate the interactions of the process parameters. In generating the CCD, the experimental domain for independent factors for high and low was entered in terms of alpha value of 2 to give 30 experiments with 6 center points. The percentage removal of Ca(II) ions was taken as the dependent variable. The Design Expert software was also

TABLE I: Experimental domain.

	pH	Contact time (min)	Initial concentration (mg L <sup>-1</sup> )	Dosage (g)
Min. point	2	10	5	0.1
Max. point	10	180	120	2

used for statistical, regression, and graphical analysis of the obtained data. The general Logit  $R$  transformation quadratic model was proposed to be

$$\begin{aligned} \text{Logit } R = & b_0 + b_1A + b_2B + b_3C + b_4D + b_5A^2 + b_6B^2 \\ & + b_7C^2 + b_8D^2 + b_9AB + b_{10}AC + b_{11}AD \quad (3) \\ & + b_{12}BC + b_{13}BD + b_{14}CD, \end{aligned}$$

where  $R$  is the percentage removal of Ca(II) ions,  $A$  is the pH,  $B$  is the contact time in min,  $C$  is the initial concentration in mg/L, and  $D$  is the dosage in g. The optimum values of the independent input variables were obtained by Excel Solver software applying the GRG function.

**2.6. Desorption Studies.** 0.2 g of the adsorbent was left in contact with 100 mL of 30 mg/L of Ca(II) for 180 min at 150 rpm. The spent adsorbent was filtered, washed, and dried at 55°C for 24 h. The dried spent adsorbent was left in contact with different concentrations of NaOH (0.1, 0.5, 1.0, and 2.0 M) for 500 min and then filtered and washed several times with deionized water. The desorbed adsorbent was used for adsorption of the Ca(II) ions and the regeneration percentage was determined.

### 3. Results and Discussion

#### 3.1. Characterization of the Adsorbent

**3.1.1. FTIR.** Zeolites were synthesized at different reaction times and different NaOH concentrations and the zeolite that was similar to zeolite-A was selected for use. Figure 1 shows the FTIR spectra of the metakaolin zeolites synthesized at an optimum time of 4 h at different NaOH concentrations. The spectrum of the selected zeolite (C2, Figure 1) had a band at 3458.53 cm<sup>-1</sup> corresponding to O-H vibration. There was structural T-O stretching and deformation at 1642.33 cm<sup>-1</sup> and a peak at 1021.10 cm<sup>-1</sup> that corresponded to internal tetrahedron asymmetrical vibration. A peak at 1421.37 cm<sup>-1</sup> might be due to Si-O asymmetrical stretching whilst bands at 723.79 and 1072.79 cm<sup>-1</sup> were due to internal symmetrical stretching vibration of (Al-Si)-O. The 473.46 cm<sup>-1</sup> band was due to internal vibration of (Al-Si)-O bending and the 558.95 cm<sup>-1</sup> band corresponded to the formation of double four-membered rings by external linkages [5]. The zeolite-A framework possesses good exchange capacity, making it a good material as an ion exchange agent. It may thus be concluded that IR spectra of the C2 spectra are closely complementary to zeolite-A; therefore, the C2 zeolites were used for batch adsorption studies.

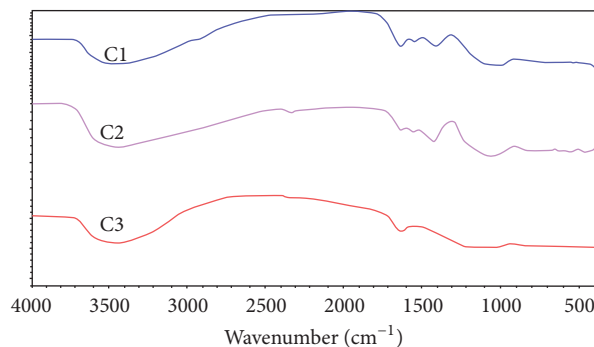


FIGURE 1: FTIR spectra of metakaolin zeolites (C1, 4 h, 3 M NaOH zeolite; C2, 4 h, 5 M NaOH zeolite; and C3, 4 h, 1 M NaOH zeolite).

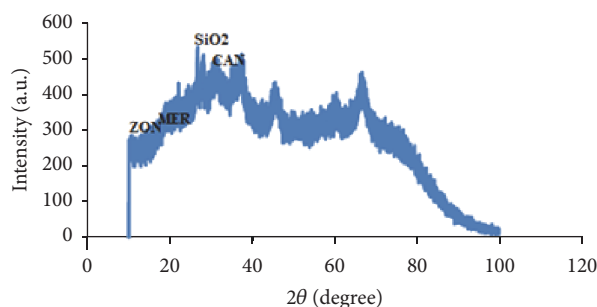


FIGURE 2: XRD pattern for metakaolin zeolite-A.

**3.1.2. XRD Analysis.** The XRD data (Figure 2) showed four crystallographic phases at  $2\theta$  degree values of 10.09 (ZON), and 12.69 (MER), 26.69 (Quartz), and 30.28 (CAN) are quite consistent with the two theta values of zeolite-A (LTA) showing that the pure phase of crystalline zeolite-A was produced.

**3.2. Adsorption Batch Studies.** Adsorption studies using MZ were done on an aqueous Ca(II) solution applying a central composite design [24] developed from the experimental domain presented in Table 1. The investigation was carried out following a full central composite design consisting of 30 experiments, which included several replicate experiments to check for consistency of the adsorption procedure. The average Ca(II) removal from aqueous solutions using the prepared adsorbent over six replicate runs (midpoint experimental conditions) was  $39.73 \pm 0.41\%$  (at 95% confidence). The results obtained were consistent as evidenced by the small relative standard deviation (1.04%). The CCD batch method was therefore capable of providing consistent results and hence was suitable for use in the investigation of the adsorption of Ca(II) ions onto MZ.

**3.2.1. Response Surface Methodology Modeling.** The experimental variable settings used for this investigation together with the actual experimental results are presented in Table 2.

The proposed full Logit  $R$  transformation model (see (3)) was fitted on the experimental data obtained in Table 2. The least squares regression results for the adsorption of Ca(II)

TABLE 2: Experimental variables and results.

Run	A: pH	B: contact time (min)	C: initial concentration (mgL <sup>-1</sup> )	D: dosage (g)	Ca(II) removal (%)
1	2	95	62.5	1.05	32.0
2	4	52.5	33.75	0.575	57.8
3	4	52.5	33.75	1.525	62.2
4	4	52.5	91.25	1.525	26.8
5	4	52.5	91.25	0.575	24.7
7	4	137.5	33.75	0.575	74.1
8	4	137.5	33.75	1.525	78.5
9	4	137.5	91.25	1.525	28.5
10	4	137.5	91.25	0.575	25.8
11	6	10	62.5	1.05	36.0
12	6	95	62.5	1.05	39.2
13	6	95	62.5	1.05	39.2
14	6	95	62.5	1.05	40.0
15	6	95	62.5	2	42.4
16	6	95	62.5	1.05	40.0
17	6	95	62.5	0.1	36.0
18	6	95	62.5	1.05	40.0
19	6	95	62.5	1.05	40.0
20	6	95	120	1.05	22.9
21	6	180	62.5	1.05	43.2
22	8	52.5	33.75	0.575	66.7
23	8	52.5	33.75	1.525	72.6
24	8	52.5	91.25	1.525	30.1
25	8	52.5	91.25	0.575	26.8
26	8	137.5	33.75	1.525	75.6
27	8	137.5	33.75	0.575	59.3
28	8	137.5	91.25	1.525	34.5
29	8	137.5	91.25	0.575	32.3
30	10	95	62.5	1.05	47.2

TABLE 3: Least squares regression data for Ca(II) removal.

Factor	Coefficient estimate	DF	Standard error	$t$ for $H_0$ Coeff. = 0	Prob > $ t $	VIF
Intercept	2.16972	1	0.2366			
A: pH	0.062119	1	0.0110	5.64648	<0.0001	1.01478
B: contact time	0.008357	1	0.0015	5.48405	<0.0001	8.79209
C: concentration	-0.0863	1	0.0065	-13.279	<0.0001	47.681
D: dosage	0.149761	1	0.0463	3.23304	0.0042	1.01478
$C^2$	0.0005	1	0.0000	10.147	<0.0001	44.4523
BC	-8.10E - 05	1	0.0000	-3.6024	0.0018	14.1953
$R^2$	0.984					

after elimination of the insignificant terms ( $p > 0.05$ ) during stepwise multiple regression are summarized in Table 3.

All the  $p$  values associated with the coefficients of pH (A), contact time (B), initial concentration (C), dosage concentration (D), initial concentration squared term ( $C^2$ ), and the

interaction between contact time and initial concentration term (BC) were smaller than 0.05 (Table 3), indicating a significant effect of these factors on the percentage removal of calcium by metakaolin zeolite. The interaction between contact time and initial concentration (BC) has an antagonistic

TABLE 4: Results for ANOVA for the model.

Source	Sum of squares	DF	Mean square	F-value	Prob > F
Model	13.4078	6	2.23463	203.981	<0.0001
Residual	0.219102	20	0.010955		
Lack of fit	0.21761	15	0.014507	48.6273	0.0002
Pure error	0.001492	5	0.000298		
Cor. total	13.6269	26			

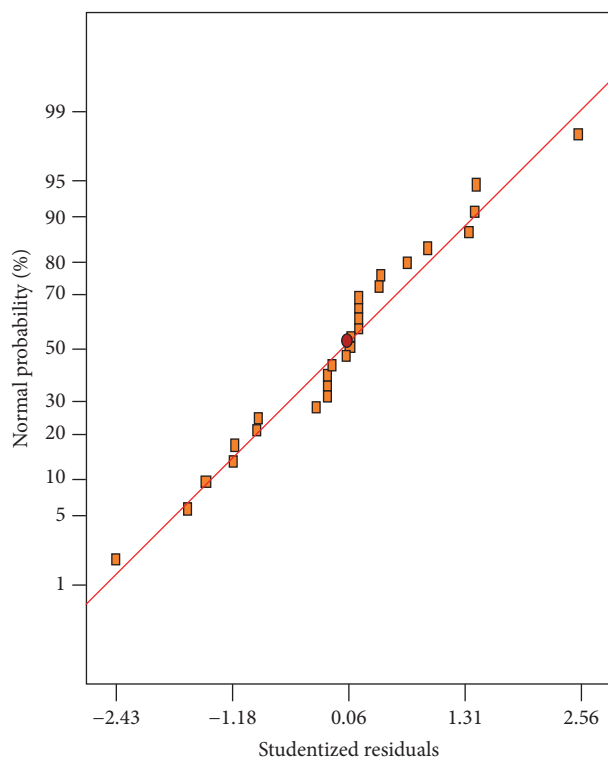


FIGURE 3: Normal plot of residuals.

effect on the removal of calcium as shown by the negative value of the associated coefficient. The final model obtained after elimination of the insignificant term was

$$\text{Logit } R = b_0 + b_1A + b_2B + b_3C + b_4D + b_5C^2 + b_6BC. \quad (4)$$

The results of the analysis of variance for the regression model are presented in Table 4.

The  $p$  value for the model was less than 0.05 ( $p < 0.0001$ ), indicating that the model was statistically significant. Moreover, a higher regression coefficient ( $R^2 = 0.984$ ) supported the significance of the model [25]. Model validity was evaluated by normal probability plot which was produced by the Design Expert software. Figure 3 shows that the residuals were close to the straight line and were scattered randomly around it with no particular pattern, showing no abnormal behavior [26].

The plot of standard residuals versus predicted values showed a random distribution of residuals (difference between the experimental and the predicted values), with no particular pattern showing well behaved residuals

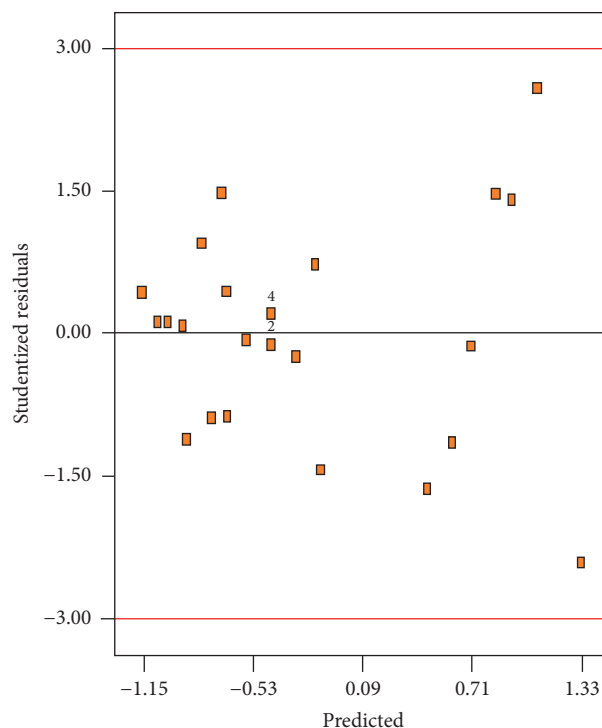


FIGURE 4: Plot of standard residuals versus predicted values.

(Figure 4) [27]. Moreover, it showed no outliers supporting the statistical validity of the model.

**3.2.2. Model Interpretation.** A 3-dimensional surface plot gives vital information on the behavior of the experimental design system. This aids the evaluation of the effects of the independent variables on the response factor. Figure 5 shows the effect of pH and contact time on % removal of Ca(II) ions at constant dosage amount and initial Ca(II) concentration. An increase in both pH and contact time increased the percentage removal of Ca(II). Since the sorbent had abundant binding sites, the gradual occupancy of these sites over different times resulted in an increase in the uptake of Ca(II). The lower percentage removal observed in acidic pH medium is due to protonation of adsorbent functional groups or competition of  $H^+$  with Ca(II) ions to bind and occupy the active sites of the adsorbent. On the other hand, higher pH medium promotes formation of insoluble metal hydroxides such as  $Ca(OH)_2$  which are filtered off.

The linear increase in adsorption efficiency of MZ with increasing adsorbent dosage indicated the accessibility of a

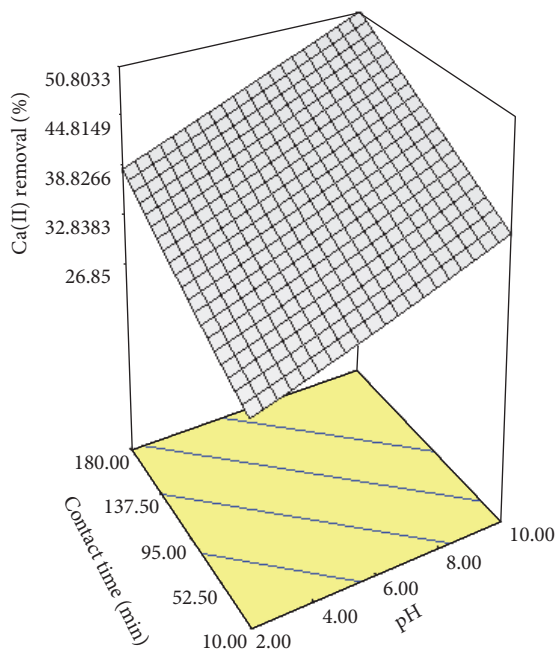


FIGURE 5: 3D surface plot for total hardness removal.

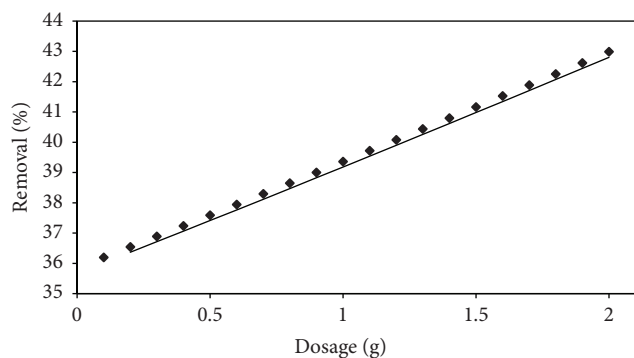


FIGURE 6: Effect of adsorbent dosage.

larger number of sorption sites at higher dosage to adsorb Ca(II) ions (Figure 6).

The effect of initial metal concentration on the removal efficiency of the zeolite adsorbent is illustrated in Figure 7. The initial concentration-profile plot shows a remarkable decrease in removal efficiency as initial concentration is increased. At low concentration, most of the metal ions in the solution interact with active sites on the adsorbent, resulting in high percentage removal [28].

**3.2.3. Effect of Temperature.** The amount of Ca(II) adsorbed on MZ as a function of solution temperature is shown in Figure 8. The removal efficiency decreased as temperature was increased from 20 to 40 °C, showing that low temperature favors Ca(II) removal from aqueous solutions. The decrease of Ca(II) removal may be attributed to the Ca(II) ions escaping from the solid phase with the rise in temperature of solutions [29]. The optimum temperature of 20 °C was used for all further experiments.

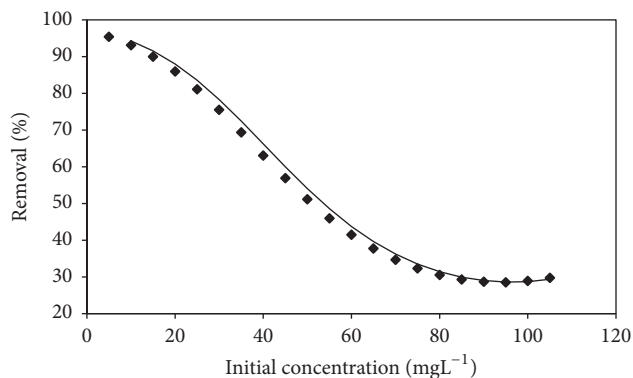


FIGURE 7: Effect of initial concentration.

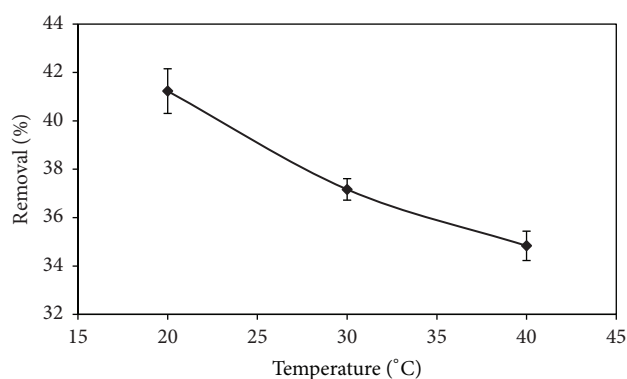


FIGURE 8: Effect of temperature.

**3.2.4. Adsorption Thermodynamics.** Thermodynamic parameters such as standard Gibbs free energy change ( $\Delta G^\circ$ ), the entropy change ( $\Delta S^\circ$ ), and the enthalpy change ( $\Delta H^\circ$ ) were calculated from the following equations:

$$\Delta G^\circ = -RT \ln K_d, \quad (5)$$

$$K_d = \frac{q_e}{C_e}$$

$$\Delta G^\circ = \Delta H^\circ - T\Delta S^\circ,$$

$$\ln K_d = \frac{\Delta S^\circ}{R} + \frac{\Delta H^\circ}{RT},$$

where  $R$  is the ideal gas constant,  $T$  is the temperature in K, and  $K_d$  is the thermodynamic equilibrium constant. The parameter  $q_e$  is the amount of Ca(II) adsorbed at equilibrium in mg/L and  $C_e$  is the equilibrium concentration of Ca(II) ions in the solution. The intercept and the slope of the Van't Hoff plots of  $\ln K_d$  versus  $1/T$  were used to determine the  $\Delta S^\circ$  and  $\Delta H^\circ$  values, respectively. Table 5 shows the thermodynamic parameters. The calculated  $\Delta H^\circ$  value was positive, showing that the adsorption process was endothermic. A slightly positive  $\Delta S^\circ$  indicated some randomness at the solid/liquid interface during the sorption process. The  $\Delta G^\circ$  values were all positive, showing that the sorption of metals was not thermodynamically spontaneous.

TABLE 5: Thermodynamic parameters.

Temperature (K)	$\Delta G^\circ$ (kJmol <sup>-1</sup> )	$\Delta H^\circ$ (kJmol <sup>-1</sup> )	$\Delta S^\circ$ (kJmol <sup>-1</sup> )
293.15	+1.284	+3.953	+0.0091
303.15	+1.193		
313.15	+1.102		

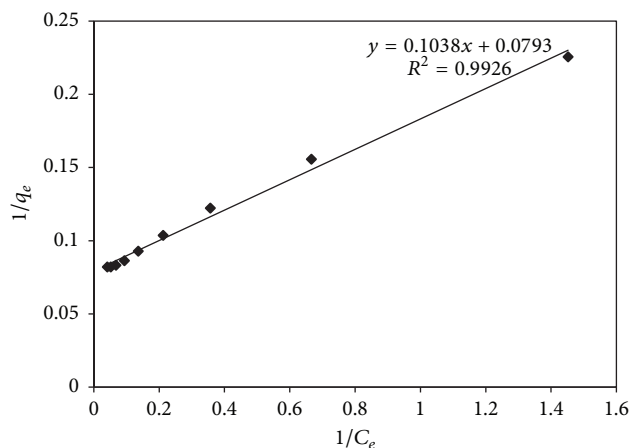


FIGURE 9: Langmuir isotherm model.

3.2.5. *Adsorption Isotherms.* Adsorption isotherms are useful in determining the nature of the interaction between the adsorbate and the adsorbent. In this study, the Freundlich [30] and Langmuir [31] adsorption isotherms were used to interpret the experimental data.

The linearized form of the Langmuir isotherm model can be represented by

$$\frac{1}{q_e} = \frac{1}{q_{\max}} + \left( \frac{1}{bq_{\max}} \right) \frac{1}{C_e}, \quad (6)$$

where  $q_e$  is the amount of metal ion adsorbed at equilibrium,  $q_{\max}$  is the maximum adsorption capacity (mg g<sup>-1</sup>),  $C_e$  (mgL<sup>-1</sup>) is the equilibrium concentration, and  $b$  is the equilibrium Langmuir constant. A plot of  $1/q_e$  versus  $1/C_e$  (Figure 9) gave a straight line with  $1/q_{\max}$  as intercept and  $1/bq_{\max}$  as slope, and hence  $q_{\max}$  and  $b$  could be estimated (Table 6).

Monolayer sorption with a heterogeneous energetic distribution of active sites, coupled by interaction between the adsorbed molecules, is assumed by the Freundlich isotherm. The linearized Freundlich isotherm equation is given by

$$\ln q_e = \ln K_F + \frac{1}{n} \ln C_e, \quad (7)$$

where  $q_e$  (mg g<sup>-1</sup>) is the amount of metal ion adsorbed at equilibrium by the adsorbent,  $K_F$  (mg g<sup>-1</sup>) and  $n$  are constants representing the adsorption capacity and intensity of adsorption, respectively, and  $C_e$  (mgL<sup>-1</sup>) is the equilibrium concentration. A plot of  $\ln q_e$  versus  $\ln C_e$  (Figure 10) gave a straight line, and  $K_F$  and  $n$  were estimated from the intercept and slope, respectively (Table 6). The calculated values of  $n$

TABLE 6: Equilibrium studies parameters.

Isotherm	Parameter	Value
Langmuir	$q_{\max}$ (mg g <sup>-1</sup> )	12.610
	$b$ (L mg <sup>-1</sup> )	0.764
	$\chi^2$	4.76
	$R^2$	0.993
Freundlich	$KF$ (mg g <sup>-1</sup> )	5.608
	$n$	3.578
	$\chi^2$	37.91
	$R^2$	0.933

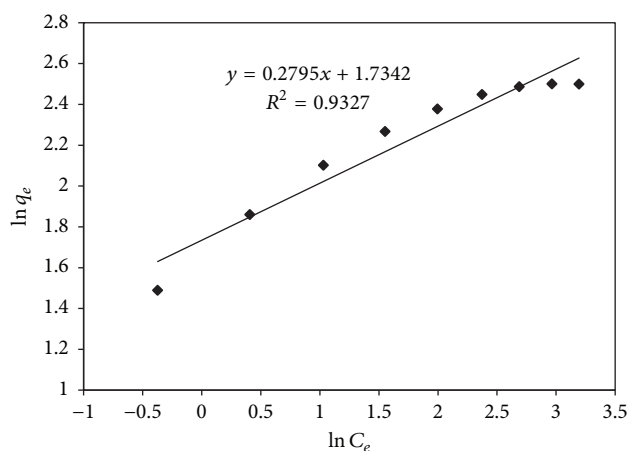


FIGURE 10: Freundlich isotherm model.

(3.578) were between 1 and 10 indicating favorability of the adsorption of Ca(II) ions on the adsorbent.

The Langmuir isotherm model correlation coefficient ( $R^2 = 0.993$ ) was significantly closer to unity relative to that of the Freundlich isotherm model ( $R^2 = 0.933$ ). Moreover, the Chi-square value for the Langmuir model ( $\chi^2 = 4.76$ ) was lower than that of Freundlich model ( $\chi^2 = 37.91$ ), indicating that the Freundlich model could not fit all the experimental data well.

3.2.6. *Adsorption Kinetics.* In the current study, the pseudo-first-order and pseudo-second-order models were used for determining the mechanism of the adsorption process. The pseudo-first-order model which assumes that the rate of adsorption site occupation is proportional to the number of unoccupied sites is expressed by the Lagergren equation [32] (Lagergren, 1898) in the linear form as follows:

$$\log(q_e - q_t) = \log q_e - \frac{t \cdot k_1}{2.3}, \quad (8)$$

where  $q_e$  and  $q_t$  are the amount of Ca(II) adsorbed (mg g<sup>-1</sup>) onto MZ at equilibrium and at any time  $t$ , respectively,  $k_1$  (min<sup>-1</sup>) is the rate constant of pseudo-first-order model, and  $t$  is the time (min). A plot of  $\log(q_e - q_t)$  versus  $t$  (Figure 11) was used to estimate the values of  $k_1$  and  $q_e$  from the slope and the intercept, respectively.

TABLE 7: Kinetic parameters.

Experimental value $q_e$ (mg g <sup>-1</sup> )	Pseudo-first-order kinetic model			Pseudo-second-order kinetic model		
	$q_e$ (mg g <sup>-1</sup> )	$k_1$ (min <sup>-1</sup> )	$R^2$	$q_e$ (mg g <sup>-1</sup> )	$k_1$ (g mg <sup>-1</sup> min <sup>-1</sup> )	$R^2$
2.77	1.01	0.01	0.964	2.81	0.0318	0.991

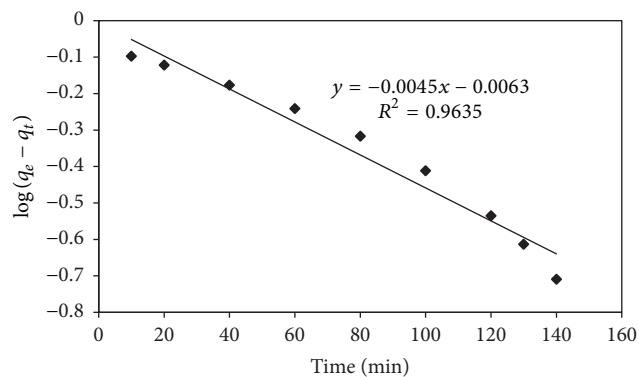


FIGURE 11: Pseudo-first-order model.

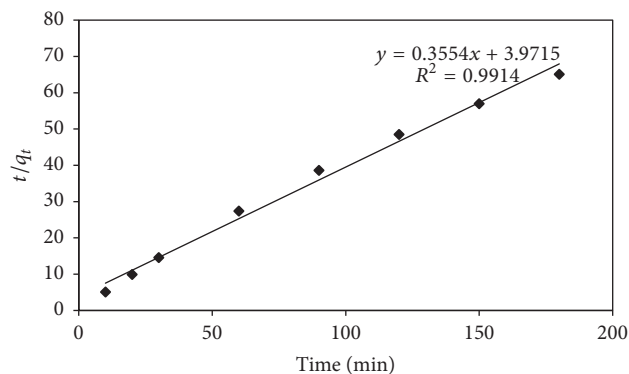


FIGURE 12: Pseudo-second-order kinetic model.

The second-order kinetic model can be expressed in the linear form as follows [33]:

$$\frac{t}{q_t} = \frac{t}{q_e} + \frac{1}{k_2 (q_e)^2}, \quad (9)$$

where  $k_2$  is the equilibrium rate constant of pseudo-second-order model (g mg<sup>-1</sup> min<sup>-1</sup>). The values of  $q_e$  and  $k_2$  (Table 7) were determined from the slope and intercept of the plot of  $(t/q_t)$  versus  $t$ , respectively (Figure 12). The pseudo-second-order kinetic model correlation coefficient was  $R^2 = 0.991$ , suggesting the applicability of the model to the adsorption of Ca(II) by MZ. Furthermore, the  $q_{e,cal}$  value (2.81 mg g<sup>-1</sup>) for the pseudo-second-order kinetic reaction was found to be in better agreement with the  $q_{e,exp}$  value (2.77 mg g<sup>-1</sup>).

**3.2.7. Desorption Studies.** The reusability of the spent adsorbent was studied by desorption/regeneration experiments, to evaluate the economical feasibility of the method. Regeneration percentage of the spent adsorbent versus NaOH is

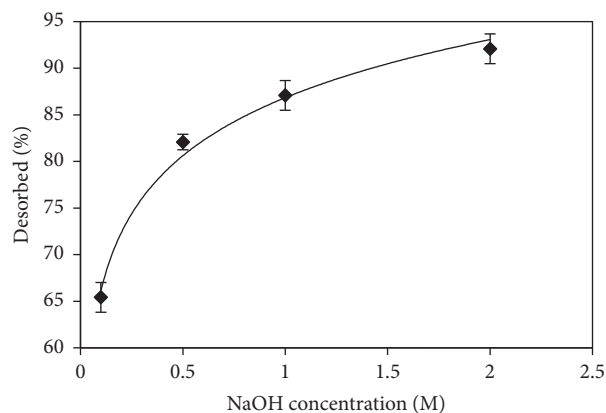


FIGURE 13: Desorption of Ca(II) ions from the spent adsorbent.

illustrated in Figure 13. Maximum recovery of the adsorbed metal ions was achieved at 2 M NaOH concentration leading to 90% desorption of the adsorbed Ca(II).

## 4. Conclusions

The present study shows that metakaolin zeolite has much potential as an adsorbent for the removal of Ca(II) from aqueous solutions. The influences of pH, adsorbent dosage, contact time, and initial concentration on adsorption of the metal ion and their interactions were investigated by a response surface methodology and ANOVA. The Excel Solver showed that optimum removal (87.70%) was obtained at a pH, contact time, adsorbent dosage, and initial Ca(II) concentration of 6, 180 min, 1.0 g, and 40.0 mg L<sup>-1</sup>, respectively. Confirmatory experimental analysis under the optimized conditions gave average Ca(II) removal of  $87.26 \pm 0.31\%$ . The adsorption process was found to be endothermic and not thermodynamically spontaneous. The Langmuir isotherm model best described the adsorption mechanism with maximum adsorption capacity of 12.61 mg g<sup>-1</sup>. The kinetic study was performed based on pseudo-first-order and pseudo-second-order equations and the adsorption followed pseudo-second-order kinetic model. The results obtained showed that metakaolin zeolite can be used as an effective adsorbent for removal of Ca(II) ions from aqueous solutions.

## Conflicts of Interest

The authors declare that there are no conflicts of interest regarding the publication of this paper.



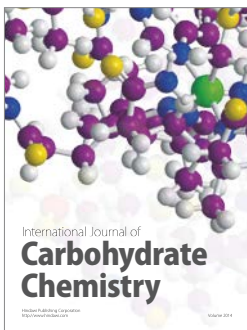
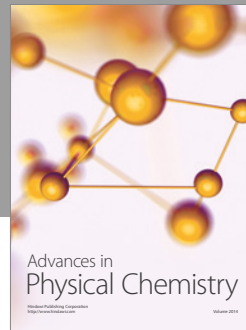
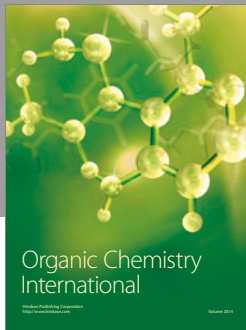
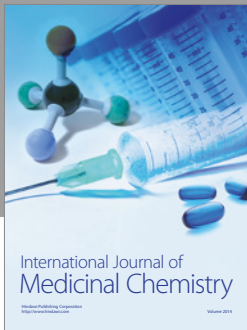
## Acknowledgments

The authors are grateful to the Department of Chemical Technology, Midlands State University, Gweru, Zimbabwe, for providing the research facilities.

## References

- [1] H. Liu, S. Peng, L. Shu, T. Chen, T. Bao, and R. L. Frost, "Magnetic zeolite NaA: synthesis, characterization based on metakaolin and its application for the removal of  $\text{Cu}^{2+}$ ,  $\text{Pb}^{2+}$ ," *Chemosphere*, vol. 91, no. 11, pp. 1539–1546, 2013.
- [2] M. Gougazeh and J.-C. Buhl, "Synthesis and characterization of zeolite A by hydrothermal transformation of natural Jordanian kaolin," *Journal of the Association of Arab Universities for Basic and Applied Sciences*, vol. 15, no. 1, pp. 35–42, 2014.
- [3] H. S. Ibrahim, T. S. Jamil, and E. Z. Hegazy, "Application of zeolite prepared from Egyptian kaolin for the removal of heavy metals: II. Isotherm models," *Journal of Hazardous Materials*, vol. 182, no. 1-3, pp. 842–847, 2010.
- [4] T. T. Walek, F. Saito, and Q. Zhang, "The effect of low solid/liquid ratio on hydrothermal synthesis of zeolites from fly ash," *Fuel*, vol. 87, no. 15-16, pp. 3194–3199, 2008.
- [5] C.-F. Wang, J.-S. Li, L.-J. Wang, and X.-Y. Sun, "Influence of NaOH concentrations on synthesis of pure-form zeolite A from fly ash using two-stage method," *Journal of Hazardous Materials*, vol. 155, no. 1-2, pp. 58–64, 2008.
- [6] T. S. Jamil, H. S. Ibrahim, I. H. Abd El-Maksoud, and S. T. El-Wakeel, "Application of zeolite prepared from Egyptian kaolin for removal of heavy metals: I. Optimum conditions," *Desalination*, vol. 258, no. 1-3, pp. 34–40, 2010.
- [7] A. A. Ismail, R. M. Mohamed, I. A. Ibrahim, G. Kini, and B. Koopman, "Synthesis, optimization and characterization of zeolite A and its ion-exchange properties," *Colloids and Surfaces A: Physicochemical and Engineering Aspects*, vol. 366, no. 1-3, pp. 80–87, 2010.
- [8] A. R. Loiola, J. C. R. A. Andrade, J. M. Sasaki, and L. R. D. da Silva, "Structural analysis of zeolite NaA synthesized by a cost-effective hydrothermal method using kaolin and its use as water softener," *Journal of Colloid and Interface Science*, vol. 367, no. 1, pp. 502–508, 2012.
- [9] M. L. Mignoni, D. I. Petkowicz, N. R. C. Fernandes Machado, and S. B. C. Pergher, "Synthesis of mordenite using kaolin as Si and Al source," *Applied Clay Science*, vol. 41, no. 1-2, pp. 99–104, 2008.
- [10] D. M. El-Mekkawi and M. M. Selim, "Removal of  $\text{Pb}^{2+}$  from water by using Na-Y zeolites prepared from Egyptian kaolins collected from different sources," *Journal of Environmental Chemical Engineering*, vol. 2, no. 1, pp. 723–730, 2014.
- [11] S. Chandrasekhar and P. N. Pramada, "Microwave assisted synthesis of zeolite A from metakaolin," *Microporous and Mesoporous Materials*, vol. 108, no. 1-3, pp. 152–161, 2008.
- [12] M. N. Sepehr, M. Zarrabi, H. Kazemian, A. Amrane, K. Yaghmaian, and H. R. Ghaffari, "Removal of hardness agents, calcium and magnesium, by natural and alkaline modified pumice stones in single and binary systems," *Applied Surface Science*, vol. 274, pp. 295–305, 2013.
- [13] A. Dimirkou and M. K. Doula, "Use of clinoptilolite and an Fe-overexchanged clinoptilolite in  $\text{Zn}^{2+}$  and  $\text{Mn}^{2+}$  removal from drinking water," *Desalination*, vol. 224, no. 1-3, pp. 280–292, 2008.
- [14] Z. Xue, Z. Li, J. Ma et al., "Effective removal of  $\text{Mg}^{2+}$  and  $\text{Ca}^{2+}$  ions by mesoporous LTA zeolite," *Desalination*, vol. 341, no. 1, pp. 10–18, 2014.
- [15] L. Fu, J. Wang, and Y. Su, "Removal of low concentrations of hardness ions from aqueous solutions using electrodeionization process," *Separation and Purification Technology*, vol. 68, no. 3, pp. 390–396, 2009.
- [16] J. N. Apell and T. H. Boyer, "Combined ion exchange treatment for removal of dissolved organic matter and hardness," *Water Research*, vol. 44, no. 8, pp. 2419–2430, 2010.
- [17] L. Seifi, A. Torabian, H. Kazemian et al., "Adsorption of BTEX on surfactant modified granulated natural zeolite nanoparticles: parameters optimizing by applying taguchi experimental design method," *Clean - Soil, Air, Water*, vol. 39, no. 10, pp. 939–948, 2011.
- [18] M. Roosta, M. Ghaedi, A. Daneshfar, R. Sahraei, and A. Asghari, "Optimization of the ultrasonic assisted removal of methylene blue by gold nanoparticles loaded on activated carbon using experimental design methodology," *Ultrasonics Sonochemistry*, vol. 21, no. 1, pp. 242–252, 2014.
- [19] M. Roosta, M. Ghaedi, A. Daneshfar, and R. Sahraei, "Experimental design based response surface methodology optimization of ultrasonic assisted adsorption of safranin O by tin sulfide nanoparticle loaded on activated carbon," *Spectrochimica Acta - Part A: Molecular and Biomolecular Spectroscopy*, vol. 122, pp. 223–231, 2014.
- [20] M. K. Satapathy and P. Das, "Optimization of crystal violet dye removal using novel soil-silver nanocomposite as nanoadsorbent using response surface methodology," *Journal of Environmental Chemical Engineering*, vol. 2, no. 1, pp. 708–714, 2014.
- [21] M. Savasari, M. Emadi, M. A. Bahmanyar, and P. Biparva, "Optimization of Cd (II) removal from aqueous solution by ascorbic acid-stabilized zero valent iron nanoparticles using response surface methodology," *Journal of Industrial and Engineering Chemistry*, vol. 21, pp. 1403–1409, 2015.
- [22] C. A. Ríos, C. D. Williams, and M. A. Fullen, "Nucleation and growth history of zeolite LTA synthesized from kaolinite by two different methods," *Applied Clay Science*, vol. 42, no. 3-4, pp. 446–454, 2009.
- [23] P. Thuadaj and N. Nuntiya, "Synthesis of Na-x hydrate zeolite from fly ash and amorphous silica from rice husk ash by fusion with caustic soda prior to incubation," in *Proceedings of the International Conference on Chemistry and Chemical Process IPCBEE*, vol. 10, pp. 69–74, 2011.
- [24] S. Sugashini and K. M. M. S. Begum, "Optimization using central composite design (CCD) for the biosorption of Cr(VI) ions by cross linked chitosan carbonized rice husk (CCACR)," *Clean Technologies and Environmental Policy*, vol. 15, no. 2, pp. 293–302, 2013.
- [25] A. Fakhri, "Application of response surface methodology to optimize the process variables for fluoride ion removal using maghemite nanoparticles," *Journal of Saudi Chemical Society*, vol. 18, no. 4, pp. 340–347, 2014.
- [26] A. C. Cameron and P. K. Trivedi, *Regression analysis of count data*, vol. 53 of *Econometric Society Monographs*, Cambridge University Press, Cambridge, England, 2nd edition, 2013.
- [27] D. C. Montgomery and E. A. Peck, *Introduction to Linear Regression Analysis*, John Wiley & Sons, New York, NY, USA, 1982.
- [28] M. Moyo, U. Guyo, G. Mawenyiyo, N. P. Zinyama, and B. C. Nyamunda, "Marula seed husk (*Sclerocarya birrea*) biomass

- as a low cost biosorbent for removal of Pb(II) and Cu(II) from aqueous solution,” *Journal of Industrial and Engineering Chemistry*, vol. 27, pp. 126–132, 2015.
- [29] U. Guyo, J. Mhonyera, and M. Moyo, “Pb(II) adsorption from aqueous solutions by raw and treated biomass of maize stover—a comparative study,” *Process Safety and Environmental Protection*, vol. 93, pp. 192–200, 2015.
- [30] H. M. Freundlich, “Over the adsorption in solution,” *Journal of Physical Chemistry*, vol. 57, pp. 384–470, 1906.
- [31] I. Langmuir, “The adsorption of gases on plane surfaces of glass, mica and platinum,” *The Journal of the American Chemical Society*, vol. 40, no. 9, pp. 1361–1403, 1918.
- [32] S. Lagergren, “About the theory of so-called adsorption of soluble substances,” *Kungliga Svenska Vetenskapsakademiens Handlingar*, vol. 24, pp. 1–39, 1898.
- [33] Y. S. Ho and G. McKay, “Pseudo-second order model for sorption processes,” *Process Biochemistry*, vol. 34, no. 5, pp. 451–465, 1999.



**Hindawi**

Submit your manuscripts at  
<https://www.hindawi.com>

